

Diversity and Structure of Miombo Woodlands in Mozambique Using a Range of Sampling Sizes

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Abstract: Forest assessments are essential to understand the tree population structure and diversity status of forests and to provide information for biodiversity recovery planning. Unfortunately, the majority of Miombo woodlands in Mozambique lack of inventory data, and consequently are often insufficient for management. This study aimed to assess the species richness, diversity and structure of Miombo woodlands using a range of sampling sizes in Mocuba district, Mozambique. Plant inventory was carried out in 128 systematically selected sample plots in 71.6 ha, which was divided into eight treatments, i.e., T1: 0.1 ha; T2: 0.25 ha; T3: 0.375 ha; T4: 0.5 ha; T5: 0.625 ha; T6: 0.75 ha; T7: 0.875 ha; T8: 1.0 ha, with 16 repetitions. All stems ≥ 10 cm diameter at breast height, and species name to evaluate the floristic composition, richness of tree species, diversity and diameter distribution were recorded. A total of 36,535 individuals were recorded, belonging 124 species, 83 genera and representing 31 botanical families. The most important species was the *Brachystegia spiciformis* Benth. and the richest botanical family was Fabaceae. The forest showed an average of 517 ± 85 trees/ha, and high species diversity and evenness. Analysis of covariance shows that the intercepts and slope of exponential function for diameter distribution are not significantly different for the eight treatments. Compared with the entire composite forest, inventory means were accurately estimated and size class distributions were well represented for plots ≤ 0.25 ha, for selecting an efficient sampling design suited to forest characteristics and the inventory's purpose.

Key words: Floristic composition, sampling units, species diversity, size class distribution, Miombo woodlands.

1. Introduction

The Miombo woodland is the most widespread deciduous woodland formation in Africa, covering an estimated 2.7 million km² in the Southern and Eastern of the continent receiving greater than 700 mm mean annual rainfall on nutrient-poor soils [1, 2]. Miombo woodland is distinguished from other African savanna. In Miombo woodland, woodland and forest are formed by the dominance of tree species in the family Fabaceae, subfamily Caesalpinioideae, particularly in the genera *Brachystegia*, *Julbernardia* and *Isoberlinia* [3, 4], while these genera are seldom found outside Miombo [5, 6]. Miombo is regarded as a vegetation type that has been used by humans since the upper

Pleistocene and maintained by man through a long history of cutting, cultivation and frequent dry season fires over the last 55,000 year [7]. In Mozambique, this woodland covers 70% of the country [8, 9], and traditionally provides a large number of ecosystem goods, hardwood timber, firewood and services [10], particularly in the form of edible species. Thus, it is economically important to livelihood [11, 12].

The composition and structure of Miombo woodland appears superficially to be relatively uniform over large regions, suggesting a broad similarity in key environmental conditions [13, 14]. However, differences in species composition, diversity and structure are more apparent at a local scale [15]. Today, the Miombo woodland is in a state of rapid change and its resources are under increasing pressure as more and more areas are modified or

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transformed [7]. Factors, such as geomorphic evolution of the landscape [16], edaphic factors, principally soil moisture and nutrients [2] and past and present land use, and other anthropogenic disturbances [6, 17], have all been implicated. According to Chidumayo [5], anthropogenic activities play a big role in the dynamics of Miombo woodlands. Knowledge of the extent, to which tree and shrub diversity have been shaped, is inadequate. It is believed to be generally driven by a complex interplay of various forces [18]. This woodland have been considered as a disclimax and dependent on fire, however, knowledge about forest successional processes and tree species ecology are poorly understood and are therefore important tasks for researchers of Africa [19, 20].

Accordingly, Kangas and Maltamo [21] and Meilby and Nord-Larsen [22] emphasized the importance of conducting studies of floristic and structural in forests in order to provide the knowledge for maintenance of biodiversity and facilitate the exploration of its products, goods and/or services from a planned and rational way, ensuring the continued flow of these resources. Such biodiversity inventories are best integrated with the timber resource inventories in order that forest management operations can be planned [23]. Unfortunately, Mozambique does not have forest assessment or inventory data for the vast majority of its Miombo woodland, and timber potential is considered to be low. It is essential for inventories to be conducted in all woodland to make available quantitative data on the structure and composition of tree species and the diameter distribution [24]. Therefore, this is one of the tools used for understanding the succession, allowing to prior review of dynamic forest conditions, enabling development of the plant community and for biodiversity recovery planning [22]. This study assessed species richness, diversity and structure of Miombo woodlands using a range of sampling sizes in Mocuba district to predict about the develop forest

sampling strategies for biodiversity recovery planning in Mozambican Miombo woodland.

2. Materials and Methods

2.1 The Study Area Description

The study site was in the Sotomane logging concession, located in Namanjavira administrative post in north part of the Zambézia province, Mozambique (Fig. 1) at latitude 16°33'58" and 16°49'22" S, longitude 36°32'57" and 36°47'39" E, covering about 40,000 ha.

Mocuba district is located in lower Zambézia and topography is characterized by an almost flat to undulating topography, with low elevation at around 200-400 m altitude above sea level, and by sandy soils with outbreaks of sandstone, calcareous conglomerates and turf soils around the vleis and in the river floodplains. The sands are underlain by sub-littoral sands which accumulate water, enabling them to support very tall trees [25]. The climate of the district is tropical sub-humid and presents two seasons: dry winter from April to October and rainy season from November to March. Total annual precipitation is 1,200 mm [24]. According to Pereira [24], the mean temperature varies from 22 °C during dry winter months to 27 °C for the wet and hot period. October is the hottest month and July is the coldest. Whirlwinds are frequent during the hot dry period.

Vegetation of Sotomane logging concession comprises a mosaic of dry deciduous Miombo woodland and dry lowland forest with a canopy cover of 40%-70%, characterized by a variety of species that is dominated by *Brachystegia* spp., *Julbernardia globiflora* plant species, three closely related genera from the Fabaceae family, subfamily Caesalpinioideae [24, 26]. Although the dominance by *Caesalpinioideae* is characteristic, their contribution to tree numbers and biomass varies extensively within and between communities [9, 10]. The dry Miombo canopy is generally less than 15 m in height and deciduous for a month or more during the dry season [3, 5].

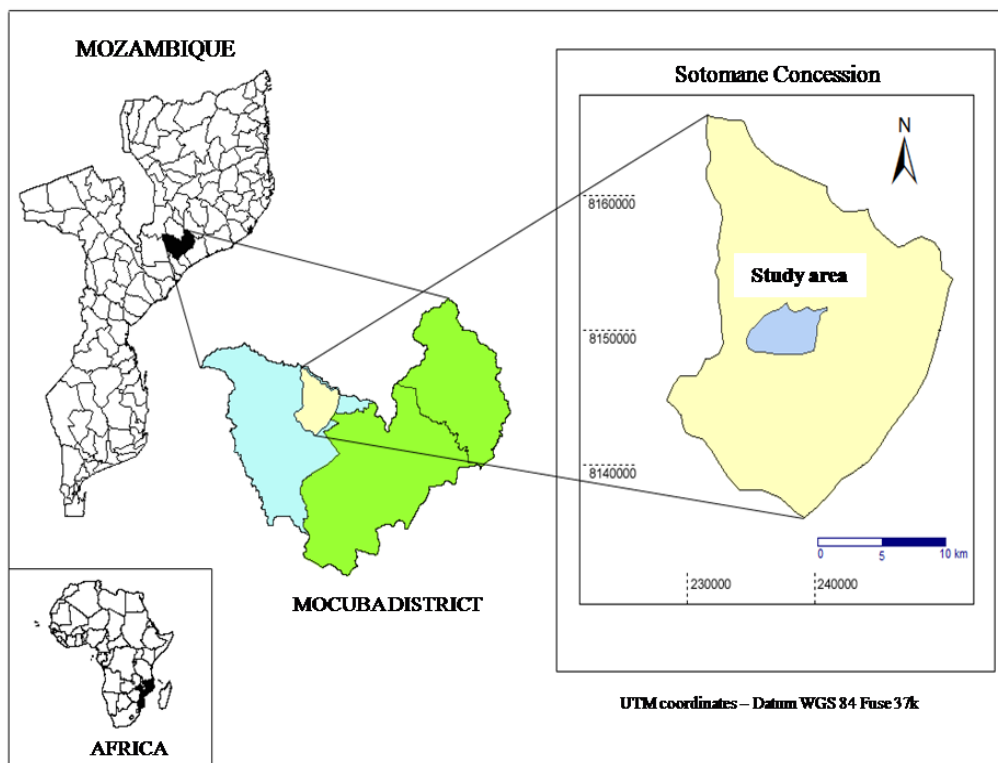


Fig. 1 The geographic location of Sotomane logging concession and the location of the Mocuba district in Mozambique.

2.2 Sampling Design and Sampling Intensity

Systematic sampling was adopted (Fig. 2). Under this design, the first sampling unit was randomly selected on the ground, and thereafter, plots were spaced at uniform intervals. The reason for the selection of this design was based on the fact that there was no great variation in the forest vegetation. According to Malimbwi and Mugasha [27], since all parts of the population are represented in the sample, the precision of the parameters estimated is usually high. Field sampling was designed to represent variations over the large spatial extent of the area, and the sampling design resulted in that no plots being placed in forest directly was affected by logging. Areas outside of this core area are intensively managed by concessionaire, and thus are not suitable for determining the natural ecological relationships [28]. The management unit is represented by a systematic sampling grid covering an area, and was preceded by a reconnaissance survey, which involved,

as sampling units, 128 fixed-area plots laying-out from the map of the forest concession. The area was gridded into 16 tracks with distant 200 m from each other, and contained eight randomly allocated units of different sizes and shape rectangular in each lane, kept a minimum distance of 300 m from each other, to cover the whole woodland area and variation between vegetation. The width of the sampling units was 50 m, which is about the limit that a man can sight into the forest of large individuals [29]. In this study, systematic sampling design ensured an even spread of the samples throughout the woodland area, and thus increase the chances of including all vegetation types. And this design was adopted, because they are easy to use in the woodland [29]. The sample plot was divided into eight treatments (sizes plots): T1: 0.1 ha; T2: 0.25 ha; T3: 0.375 ha; T4: 0.5 ha; T5: 0.625 ha; T6: 0.75 ha; T7: 0.875 ha; T8: 1.0 ha, respectively, a total of 71.6 ha. The length of the tracks followed the north-south direction and width from east to west using a compass [30].

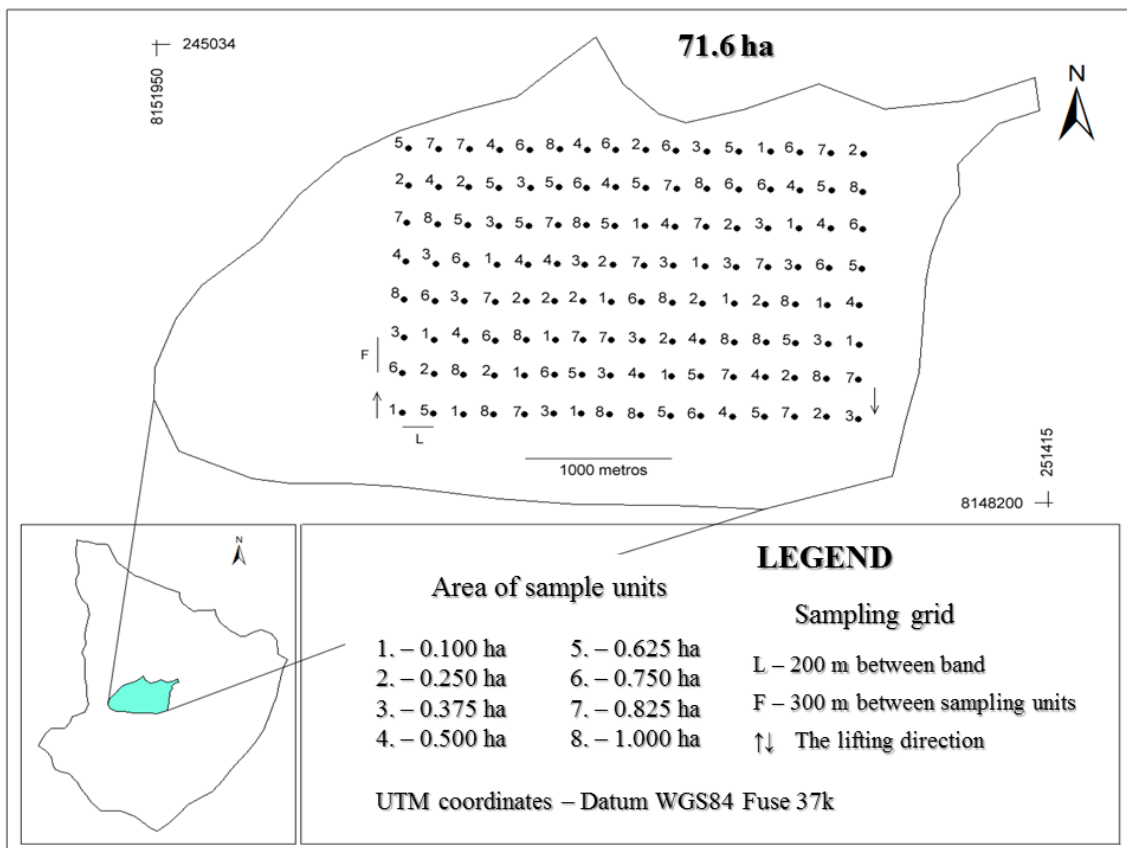


Fig. 2 Structure of the study plots system.

2.3 Field Measurements

Traditionally, in Mozambique plot, wise inventory has been used to derive the estimates for forest type level characterization. The plots were established during the dry season from June to October. And at the time of establishment, all trees/stems within the plots with diameter at breast height (dbh) ≥ 10 cm were measured, identified to species level, enumerated and numbered using GPS. Trees forking below 1.3 m were measured and recorded, separately, and at the point, dbh was measured and marked with permanent paint for subsequent re-measurements [31]. Tree species were recorded with the help of traditional botanists and one book guides of “checklist for vernacular plant names in Mozambique” [32]. Any species not identified in the field were taken as voucher specimens for identification at the Eduardo Mondlane University Research Herbarium. The

scientific names of the plants are based mainly on the angiosperm phylogeny group III [33, 34].

2.4 Data Analysis

2.4.1 Species Composition and Richness

The species richness (S) is the number of species recorded in the whole stand. The data were analyzed for species composition, richness, diversity and species importance value index (IVI). To achieve a nearly complete description of species composition and distribution, IVI was used in this study. IVI were computed as the average of the relative basal area, density and frequency. The IVI for a species is a composite of three ecological parameters—relative abundance (density of plants), relative frequency (how often a species occurs in the plots) and relative dominance (density of stock expressed as basal area), which measure different features and characteristics of a species in its habitat [35]. Thus, the IVI provides a

summary of all three indicators of ecosystem importance and is a frequently used quantifier for vegetation studies. Ecologically, the density and the frequency of a species measure the distribution of a species within the population, while the basal area measures the area occupied by the stems of trees [6].

2.4.2 Species Diversity and Evenness Index

Species diversity was computed using Shannon's diversity indexes. The Shannon's diversity index (H) is computed using the following Eq. (1):

$$H = -\sum_{i=1}^S \left(\frac{n_i}{n} \right) \times \ln \left(\frac{n_i}{n} \right) \quad (1)$$

where, n_i is the number of individuals of species i , and n is the overall number of trees inventoried in the plot.

The knowledge of species diversity is useful for establishing the influence of biotic disturbance and the state of succession and stability in the environment [36, 37]. This species diversity index increases with the number of species in the community [6, 38].

The Pielou's evenness (Eq) measures the diversity degree of a stand compared with the possible maximum and is computed as Eq. (2):

$$Eq = \frac{H}{H_{\max}} \quad (2)$$

$$H_{\max} = \ln(S)$$

where, H_{\max} is the maximum value of the Shannon's diversity index of the stand, and S is the number of tree-species recorded in the considered plot [36, 38, 39].

All diversity and evenness indices were calculated using PC-ORD for Windows version 6.0 [40]. The significance of the influence of the size of the plots on differences of Shannon's diversity index was tested using analysis of variance (ANOVA), and the mean number of individual per species was compared among sample size followed by a post hoc Tukey test.

2.4.3 Diameter Distribution

To assess the structure of the forest, the diameter was adjusted according to the Meyer function [20], for a class of individuals dbh for the population and for the eight treatments analyzed separately, using the

following Eq. (3):

$$N_i = e^{(\beta_0 + \beta_1 \cdot cc_i)} \quad (3)$$

where, N_i and cc_i are the number of stems and midpoint of the i th diameter class, respectively, β_0 and β_1 are the parameters of the distribution. The parameter β_1 determines the rate at which trees diminish in successive diameter classes, and β_0 indicates the relative density of the stand [20].

The slopes and elevation were compared between the levels and frequency distributions by class of dbh in the eight treatments assessed by applying ANCOVA [39] to the straight lines obtained by the standard regression method. To determine the differences between treatments and frequency distributions by class of dbh characteristics, data and Meyer function were ln-transformed into a linear model for homogenizing variances with additive error, as Eq. (4):

$$\ln(N_i) = \beta_0 + \beta_1 \cdot cc_i \quad (4)$$

where, $\ln(N_i)$ is the natural logarithm of number of individual by diametric class; β_0 and β_1 are estimated regression coefficient; cc_i is class center; T_i —treatment; $(T_i \times cc_i)$ —interactions.

The adhesion test Chi-square (χ^2) was applied to verify the degree of agreement of the frequency distribution of class dbh observed following model proposed by Meyer set [41]. Data analysis was carried out in the Statistical Analysis System 9.2 (SAS Institute Inc., Cary, NC). All the data were considered for each treatment separately as a single entity, or pooled without distinguishing among forest types at significant level $P < 0.05$.

3. Results and Discussion

3.1 Species Composition and Richness

A total of 36,535 stems were identified (Table 1). These represent 31 botanic families and 83 genera. Fabaceae is the largest family with 48 species, followed by Euphorbiaceae with 11 species. None of the other families have more than 10 species.

Table 1 Tree species composition and richness in the Mocuba district, ranked by IVI.

Rank	Tree species	Number of individual specie	Relative density (%)	Relative basal area (%)	Relative frequency (%)	IVI (%)
1	<i>Brachystegia spiciformis</i> Benth.	4,020	11.0	11.1	1.9	8.0
2	<i>Pterocarpus angolensis</i> DC.	2,381	6.5	6.3	1.9	4.9
3	<i>Pseudolachnostylis maprouneifolia</i> Pax	1,791	4.9	4.9	1.9	3.9
4	<i>Swartzia madagascariensis</i> Desv.	1,668	4.6	4.5	1.9	3.7
5	<i>Julbernardia globiflora</i> (Benth.) Troupin	1,248	3.4	3.3	1.8	2.9
6	<i>Burkea africana</i> Hook.	1,132	3.1	3.1	1.7	2.6
7	<i>Pteleopsis myrtifolia</i> Engl. & Diels	1,032	2.8	2.9	1.7	2.5
8	<i>Brachystegia boehmii</i> Taub.	873	2.4	2.4	1.8	2.2
9	<i>Ficus ingens</i> Miq.	831	2.3	2.3	1.3	2.0
10	<i>Pericopsis angolensis</i> (Baker) Meeuwen	743	2.0	2.0	1.8	1.9
11	<i>Azelia quanzensis</i> Welw.	744	2.0	2.0	1.8	1.9
12	<i>Cordyla africana</i> Lour.	798	2.2	2.3	1.2	1.9
13	<i>Erythrophleum suaveolens</i> (Guill. & Perr.) Brenan	734	2.0	2.0	1.6	1.9
14	<i>Dalbergia melanoxylon</i> Guill. & Perr.	677	1.9	1.7	1.8	1.8
15	<i>Combretum imberbe</i> Wawra	624	1.7	1.8	1.6	1.7
16	<i>Millettia stuhlmannii</i> Taub.	558	1.5	1.6	1.7	1.6
17	<i>Brachystegia bussei</i> Harms	610	1.7	1.7	1.1	1.5
18	<i>Parinari curatellifolia</i> Planch. ex Benth.	539	1.5	1.4	1.5	1.5
19	<i>Kirkia acuminata</i> Oliv.	509	1.4	1.4	1.5	1.5
20	<i>Pseudobersama mossambicensis</i> (Sim) Verdc.	422	1.2	1.3	1.6	1.4
Others		14,601	38.9	40.0	-	36.9
Total		36,535	100	100	100	100

The species richness (124 species) of trees and the densities of species from the Fabaceae family observed in this study compare well with the Miombo woodland occurring in other areas. In Mozambique, Ribeiro et al. [10] enumerated 79 species in Niassa Forest Reserve, while Williams et al. [9] found 69 species around Gorongosa National Park. In Kitulanhalo Forest Reserve and Ihombwa village, Backéus et al. [1] and Mbwambo et al. [42] found 79 and 86 species, respectively, from Tanzania's Miombo woodland. The average tree density of 517 ± 85 trees/ha in Miombo woodland is comparable with those reported by Backéus et al. [1], Campbell et al. [2], Kalaba et al. [4] and Chidumayo [5].

The 20 most IVI of occurring tree species are summarized as descending order in Table 1. In terms of IVI, the most important species in mature woodland in this area are *Brachystegia spiciformis* Benth., *Pterocarpus angolensis* DC., *Pseudolachnostylis*

maprouneifolia Pax var. *Maprouneifolia*, *Swartzia madagascariensis* Desv., *Julbernardia globiflora* (Benth.) Troupin, *Burkea africana* Hook., *Pteleopsis myrtifolia* Engl. & Diels., *Brachystegia boehmii* Taub., *Ficus ingens* Miq. and *Pericopsis angolensis* (Baker) Meeuwen. They together compose 34% of the total IVI. These species are typical of the dry Miombo systems of this eco-region [6, 43, 44]. The results also show that the study site is dominated by 20 species and they represent 53% of all the trees. The dominance by a several species is an indication of a high diverse forest site in terms of tree species.

It was also found that adult population of *Dalbergia melanoxylon* Guill. & Perr., *Millettia stuhlmannii* Taub. and other economic tree species were in low densities (average of 1 tree/ha), and that lack of sapling recruitment to adult population was an increasing problem. Palgrave et al. [26] found that *D. melanoxylon*, like many other timber species, is in

only a few locations in Mozambique in very small numbers per ha. In fact, throughout Africa, the densities of these economically important tree species, such as *D. melanoxylon*, are low occurring in isolated forest plots, and their existence is owing to traditional conservation practices [18, 29]. Worthwhile to mention is the presence of *Combretum* species, which are not so typical of undisturbed Miombo but tend to occupy more perturbed areas [10, 28, 35, 45].

3.2 Species Diversity and Evenness Index

The study revealed that the species richness ranged from 93 to 124 trees, and Shannon-Wiener index of diversity (H) and the Pielou index (Eq) ranged from 3.03 to 3.64 and 0.81 to 0.88, respectively, among treatments over a total survey area of 71.6 ha (Table 2). An ecosystem with H value greater than 2.0 has been regarded as medium to high diverse in terms of species [6].

Thus, study results further showed a higher diversity than other studies in the Miombo region, such as in Mozambique where Shannon indexes of 1.25 were obtained [9], and 1.05 from Tanzania's Miombo by Shirima et al. [46], but similar to diversity (3.40, 2.90 and 3.10, respectively) in the landscapes of the Handeni Hill Forest Reserve, in Tanzania [47]. The H value (3.03-3.64) from the present study is probably due to previous disturbances, which opened the canopy and enabled regeneration of light tolerant genus, like *Combretum* [10]. Likewise, the existing management practiced has probably made regenerating species to come up vigorously.

The results also showed that there were significant

differences by the F test for the diversity of Shannon index between the treatment, and by the *Tukey* test, it was observed that T_1 (0.1 ha) was significantly differed from the others ($F = 9.89$; $P < 0.01$). This suggests that although the high floristic diversity of the study area was observed, but sampling unit size of T_1 (0.1 ha) did not capture the floristic representativeness. It should be noted that the size of sample unit of T_6 (0.75 ha) showed no statistically significant differences for the diversity compared to other sizes of sampling units (treatments). This fact is explained by the presence of rocky outcrop in some sample units of such treatments. And the fact confirmed within the study area limited the occurrence of some species, like *Khaya anthotheca*, *Pseudobersama mossambicensis* and *Entandrophragma caudatum*. On the other hand, Pereira [24] observed that the history of human disturbance in the past caused the change in floristic composition and diversity in different parts of the forests in Mocuba district. Previous disturbances from bush fire, over harvesting and grazing opened the woodland canopy and gave the way for more regeneration in the woodland [6, 21]. These results would help to understand more about the impact of human disturbance on biodiversity [7].

3.3 Diameter Distributions

The observed distribution of tree species follows an inverse J -shaped trend for the eight treatments (Fig. 3). Tree frequency decreases with the increasing diameter, which is common for natural forests with active regeneration and recruitment for the eight treatments analyzed [46, 48].

Table 2 The general characteristics of tree and resources in the Mocuba district.

Parameter	Size of sampling unit (ha)							
	0.1	0.25	0.375	0.5	0.625	0.75	0.875	1.0
Richness (total number of species)	103	118	119	93	96	108	103	124
Shannon index (H)	3.03 ^a	3.50 ^b	3.41 ^b	3.55 ^b	3.64 ^b	3.43 ^b	3.57 ^b	3.59 ^b
Pielou index (Eq)	0.86	0.85	0.82	0.88	0.87	0.83	0.86	0.81
Density (stems/ha)	525	519	542	518	583	432	512	517
Stems sampled	840	2,075	3,252	4,140	5,826	5,178	6,958	8,266

^{a, b} Same letters do not differ statistically among themselves at the significance level of 5% according to the *Tukey*'s test.

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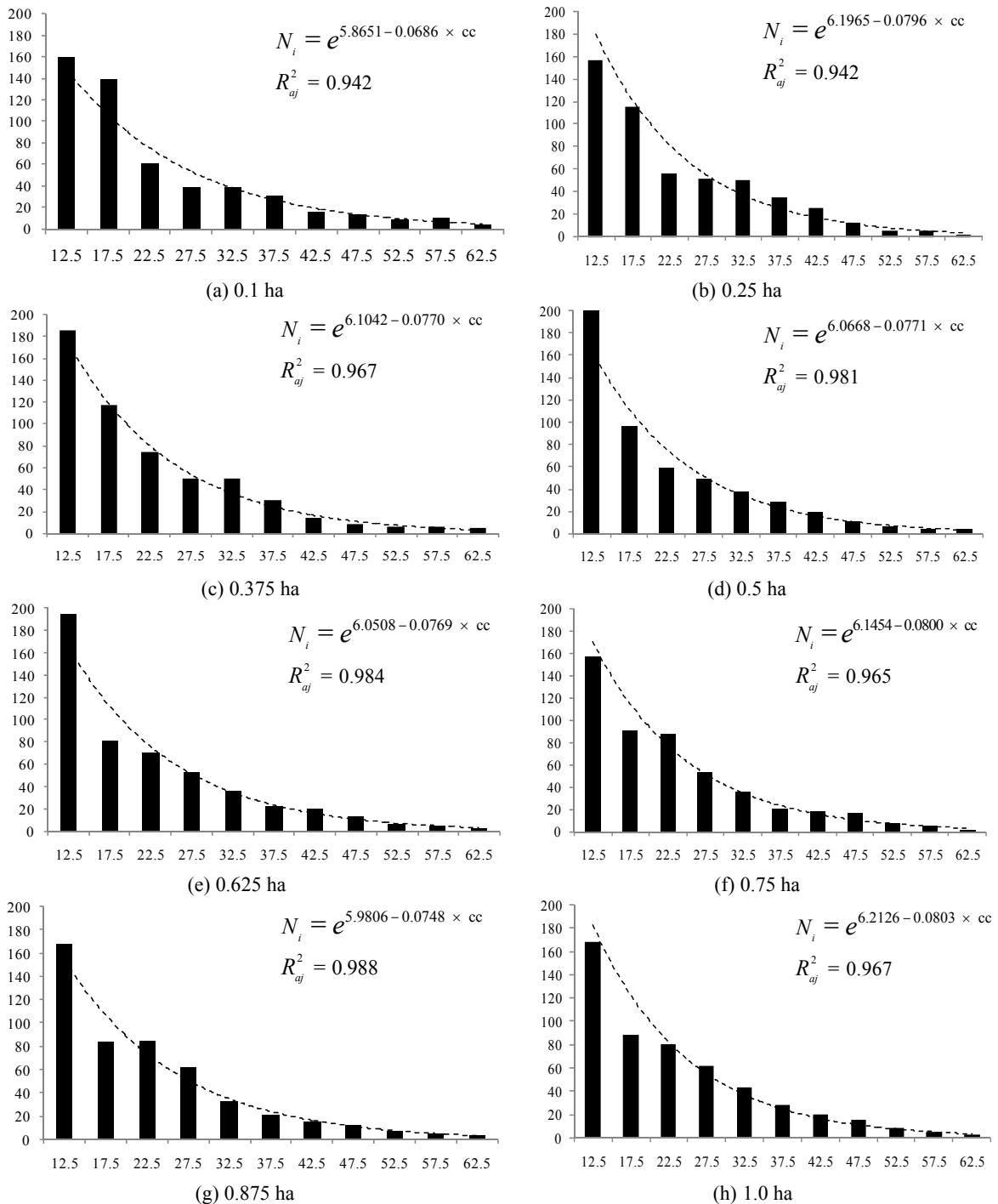


Fig. 3 Observed and predicted relative diameter distributions of the eight treatments.

cc: class center.

The histograms represent the observed distributions by diameter class (sample plots = 128). The short dashes show the distribution predicted by the negative exponential Meyer function.

In study area, there are different diameter distributions in different treatments, and human disturbances were found in all treatments, which were probably linked with illegal activities and bush fire.

All activities carried out in the woodland were under the supervision of village natural resource committees. However, absolute protection from illegal activities seems to be difficult, because of big demand of the

wood land products for their livelihoods. According to Norfolk et al. [25], few illegal activities still exist and usually are caused by people coming from non-participating villages known to have problem of firewood and charcoal. Disturbances in Miombo woodlands, such as harvesting of trees by humans for fuel and building materials are likely to greatly change the size class distribution of the harvested, may manipulate the composition and structure of tree species diversity species and also result in increased mortality [18]. Active regeneration and recruitment in Mocuba area as portrayed in this study is a good sign of sustainability of the woodland stock, which has chances of insuring sustainable supply of products and services and hence sustaining livelihoods of the woodland dependents [19]. However, Miombo woodland trees colonize faster and densely after disturbances. This happens due to the exposure of forest floor to sunlight and reduced competition among woody plants [5, 49].

On the other hand, the adhesion test χ^2 average of diametric distribution was below the critical value at the critical level of $\alpha = 5\%$, except T_1 ($\chi^2 = 19.95$; $df = 10$) and T_2 ($\chi^2 = 27.90$; $df = 10$), which showed the value of χ^2 tabulated. This indicates non-rejection of the null hypothesis for the treatments that the average value of χ^2 is below the critical value. It is concluded that the observed and expected frequencies are similar in a statistical point of view and were adherent to $\alpha = 5\%$. In this area, trees of larger diameter class ($d > 20$ cm) were observed to be less dominant in the

woodland, and this diameter class most probably suffered from the past degradation and disturbances from anthropogenic activities [9]. Trees of larger diameter classes are more suitable for commercial timber extractions, charcoal marking and firewood or construction materials [6]. This diameter class seems to be over exploited in the past with few remaining trees or limited recruitment from lower diameter classes.

From the results in Table 3, it is observed that all treatments presented for the ANCOVA showed no significant probability for interaction of $T_i \times cc_i$ ($P = 0.6137$), indicating that there are no differences between slopes in the frequency distribution by diameter class for the eight treatments. No significance was also observed in T_i ($P = 0.8741$), showing that there are no differences between levels in the frequency distribution by class of ddb in the eight treatments. There were not any pairs of treatment to be found significant differences in slopes and elevations.

Results show that the eight treatments assessed not different regressions based on levels and slopes of the curves of frequency distribution by class of dbh. This allows the use of the same function to describe the frequency distribution by class of dbh to eight treatments [50]. In Tanzania, Mwakalukwa et al. [51] achieved similar results by adjusting Meyer function using 46 sample units of different sizes, which had no statistical difference in level and inclination regardless of size variation.

Table 3 Effects of size plot (treatments) and diametric distribution on number of individual by diametric class in Miombo woodland at the study sites assessed by analysis of covariance.

Source of variation	Degrees of freedom	Sum of squares	Mean squares	F value	Probability
Model	15	130.1472	8.6764	167.37	< 0.0001
cc_i	1	129.7082	129.7082	2502.09	< 0.0001
$T_i \times cc_i$	7	0.2796	0.0399	0.77	0.6137
T_i	7	0.1595	0.0228	0.44	0.8741
Error	72	3.7325	0.0518		
Total	87	133.8797			

cc_i : class center; T_i : treatment; $T_i \times cc_i$: interactions.

4. Conclusions

The study area revealed that there is a reasonably good tree species composition and richness. Species noted to be both dominant and with high species diversity indices include *B. spiciformis* (8.0), *P. angolensis* (4.9), *P. maprouneifolia* (3.9), *S. madagascariensis* (3.7) and *J. globiflora* (2.9). These dominant as well as highly diverse tree species fit quite well within the general definition of Miombo woodland. However, it was observed that the stand is characterized by the concentration of a lot of individuals in a few species and plant families, as well as a large number of locally rare species, such as *D. melanoxylon*, *M. stuhlmanni* and *K. anthotheca*. This may be related to activity logging, being linked to management practices adopted. Generally, the number of individuals in the first diametric classes was higher in eight treatments. It indicates a good sign of sustainability of the woodland stock, which has chances of ensuring sustainable supply of products and services and hence sustains livelihoods of the surrounding communities. For selecting an efficient sampling design, suited to forest characteristics and the inventory's purpose were well represented for plots ≤ 0.25 ha.

The structural variations between plots samples showed significant differences in the diversity of species. Additionally, there was significant variability in the structural patterns in terms of diametric distribution by diameter class between treatments analyzed. The diameter structure of the forest management unit analyzed showed no differences using regression and slope-fitting techniques. This suggests sustainability of the production volume in the area, which is characterized by a low density of exploration. The results of this study apply to various spatial inventory levels and represent the most important woodland types in Mozambique. Further research is required with respect to different woodland types in other parts of the country.

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